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Expected impacts of active mining on the distribution of heavy metals in soils around Iludun-Oro and its environs, Southwestern Nigeria

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Abstract: This present study is aimed at evaluating the level of environmental degradation and potential ecological risk in soils around an active (tantalum-niobium) mining site of Iludun-Oro, southwestern Nigeria. A total of 18 topsoil samples were randomly collected from different parts of the study area and analysed using the inductively coupled plasma-mass spectrophotometry. The concentrations of Pb, Zn, Fe, Cu, Cr and Ni varied between 0–678, 7–222, 1.64–8.87, 8.8–145, 1.4–81 and 1.6–93.1 ppm, respectively, with an average concentration of 107.72, 48.17, 5.17, 45.28, 14.76 and 21.53 ppm, respectively. The order of abundance of these heavy metals analysed in the soils is as follows: Pb > Zn> Cu> Ni> Cr> Fe. The concentration values of Pb, Fe, Cu, Ba and Sr are higher than the background values indicating that mining activities have an adverse impact on the study area.

On accessing the soil quality in the study area, the results of the geo-accumulation, contamination factor (CF) and pollution load indices (PLIs) indicated high concentrations of Pb in the soils. Likewise, most of the Igeo values reported uncontaminated to moderately contaminated for all elements except for Pb. This study has further provided the background information for a vulnerability assessment of the study area as well as for future impact assessment and proper planning

ABOUT THE AUTHOR

The authors are a group of multidisciplinary scientific researchers, actively involved in the study of the ecological and geochemical study of the environments, most importantly the distribution and monitoring of elevated levels of heavy metals in soils, rocks, waters and vegetation. Their current study focuses on the assessment of heavy metals around an active mining site, its potential effects on human health using multiple indices. It provides background information for a vulnerability evaluation of the study area and recommends remediation plans for the control.

PUBLIC INTEREST STATEMENT

The impacts of mining activities through the expulsion of mine wastes pose a serious threat to human life, vegetation and animals across the globe. This perspective research describes some of the distribution of these heavy metals, their pollution level, ecological threat and potential sources of these hazardous metals around an active mining site in southwestern Nigeria. It was found that these metals range from moderately contaminated to contaminated for most of the elements. The presence of these toxic elements has adverse effects on the immediate environment. Understanding these effects can provide useful information for appropriate environmental laws and mitigate the influence of pollution on the environment in this study area and other similar places.





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of the mining site. These findings have provided useful information for making appropriate environmental management strategies to mitigate the influence of heavy metal pollution generated from an active mining site in Nigeria and other similar areas.

Subjects: Environment & Agriculture; Earth Sciences; Environmental Studies & Management; Physical Sciences

Keywords: ecological risk; heavy metals; PCA; Southwestern Nigeria; topsoil

1. Introduction

The discoveries and exploitation of natural and mineral resources in Nigeria and Africa as a whole have led to the development of the most sectors of their economy and their people's livelihood (Boakye, Dessus, Foday, & Oppong, 2012; Fayiga, Ipinmoroti, & Chirenje, 2017; Wilson, Wang, Kabenge, & Qi, 2017). The exploitation of these resources (gold, copper, tin, tantalum, columbite, niobium, natural oil and gas) has negatively impacted the environment. Several factors such as population growth, urbanization, mining, agricultural activities, and industrialization have contributed to the present pollution in the world. These global trends in environmental pollution possess a challenge, but the levels of these contaminations vary in different parts of the world, especially in developing economies.

The environmental impacts of mining activities have similar effects worldwide. These effects on the environment are determined by the chemical attributes of the ore, the method of extraction and the environmental conditions (Adriano, 2001; Higueras et al., 2004; Hsu et al., 2016; Khan, Yusoff, Bakar, Bakar, & Alias, 2016; Lupankwa, Love, Mapani, Mseka, & Meck, 2006; Mierzwa-Hersztek, Gondek, Klimkowicz-Pawlas, Baran, & Bajda, 2018; Oelofse, 2009; Potra, Dodd, & Ruhl, 2017; Yan, Liu, Zhong, Guo, & Wu, 2018). The mobility and bioavailability of these elevated heavy metals in the environment primarily depend on some physicochemical characteristics of the depositional environment such as pH, textural characteristics, organic matter content, speciation or chemical form, and electrical conductivity (Adewuyi & Osobamiro, 2016; Chukwu, 2011; Equeenuddin, Tripathy, Sahoo, & Panigrahi, 2013; Lu, Liu, Wang, & Wu, 2017; Sarapulova, Dampilova, Bardamova, Doroshkevich, & Smirnova, 2017; Zhang, You, & Hu, 2016). Higher concentrations of heavy metals in soils have been documented at a variety of sites associated with Pb-Zn mining in the southeast Missouri, USA (Potra et al., 2017). These have adverse effects on the floristic quality where no other human activities are obvious. Trace elements' (As, Cu, Zn and Pb) concentration has also been documented around a gold mine in Ghana (Antwi-Agyei, Hogarh, & Foli, 2009) with As concentrations in soils significantly higher than the Netherlands intervention value of 55 maka⁻¹. An anomalous abundance of rare earth elements in surface soils has also been reported in Bayan Obo Mine that was influenced by the exploitation of these minerals around the mining site (Wang & Liang, 2016).

The natural soil is not only a geochemical habitat for the contaminants, but also serves as a natural buffer for transportation of elements in the air, water, and biomass (Benhaddya & Hadjel, 2014; Nsenga Kumwimba et al., 2017; Wang, Deng, Yin, & Tang, 2018). Due to the high elevated values of heavy metals in soils, they have posed an adverse effect on human health and aquatic environment, which can be easily inhaled into human bodies from suspended dust or by uptake in plants (Benson, Anake, Adedapo, Fred-Ahmadu, & Ayejuyo, 2017; Jan et al., 2015; Madrid, Díaz-Barrientos, & Madrid, 2002; Mortazavi, Mortazavi, & Paknahad, 2016; Wang & Yong, 2005). Mining activities, as the hazardous sources of toxic elements, relatively affect localized areas. The tailings, gangue minerals and subsequent smelting operations have caused noticeably significant environmental concerns (Mwesigye, Young, Bailey, & Tumwebaze, 2016; Okang'Odumo et al., 2014). The fine-grained particles of mine tailings or the ore may be blown or diffused into adjacent areas. Acid-mine drainage contamination has been reported in the Abuja leather mine in Komu, southwest Nigeria (Oyebamiji, Odebunmi, Ruizhong, & Rasool, 2017), which has led to the contamination of nearby aquatic habitats.

Previous studies on the distribution of these heavy metals in our immediate environments have not only become a local phenomenon, but also a global one as well (Harikrishnan et al., 2017). Hence, geochemical studies of the mining sites using sample media like soils, stream sediments and waters are fundamental in determining the dispersion pattern of these trace elements.

In order to better understand the pollution level, ecological risk level and potential sources of heavy metals in soils of Iludun-Oro mining district, the main objectives of this study were to: (i) determine the spatial distribution of heavy metals in soils; (ii) identify their sources by using the principal component analysis (PCA); (iii) assess the contamination level using various risk assessment indices such as; CF, geoaccumulation index (Igeo), degree of contamination (CD), PLI, ecological risk index (ERI) and potential ecological risk index (PERI).

2. Materials and methods

2.1. Study area

The Iludun-Oro mining district is located in the southwestern part of Nigeria, lies between longitude 08°09'00"N to 08°16'00"N and latitude 004°49'00"E to 004°54'00"E and extends 12.5 km from east to west and 15.8 km from south to north. The study area is surrounded by several towns namely Okerimi-Oro, Agberun, Ajasse, Sanmora, Agbe-Ola, Ibode-Oro and Igbonla (Figure 1). This district is well accessible with the presence of major and minor roads created by miners and farmers during their quest to find means of livelihood. The study area is located within the tropical rainforest belt characterized by an alternation of dry and wet seasons. The rainy season runs from April through October when the effect of the southwest monsoon wind overwhelms that of the northeast trade wind, which commonly causes dry seasons from November to April. The wet season is characterized by heavy rainfall with

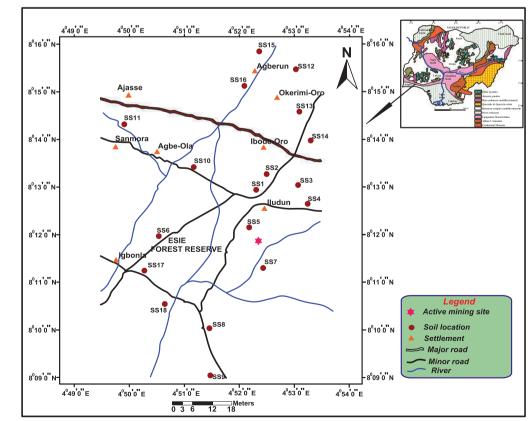


Figure 1. Location map of the study area showing sampling points.

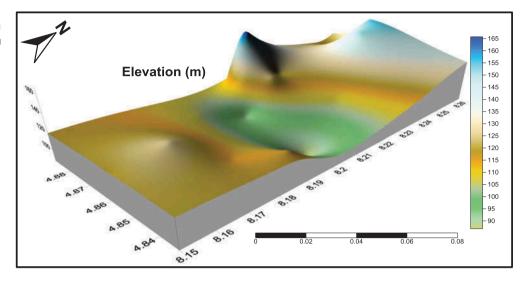


Figure 2. Geomorphological view of the study area showing the peaks and nadirs above sea level.

> corresponding low temperatures whilst the dry season comes with high temperature and little or no rain. Annual rainfall varies from 1150 mm to 2000 mm in the northern and southern areas, respectively. Annual mean temperature ranges from 21 to 29°C whilst the relative humidity ranges from 27.5 to 98.2% (Ogunrayi, Akinseye, Goldberg, & Bernhofer, 2016). The soil types mainly include reddish-brown and brown-black loam. Land use is predominantly controlled by agriculture and mining activities. The elevation of the study area varies from 90 to 165 m above sea level as shown in Figure 2. The peak of the elevation is located on the north-west flank of the study area. The area is drained by several tributaries of river running in the north-south direction. The geology of the study area is underlain by the Precambrian basement complex rocks, younger granite of Jurassic age, and the Cretaceous to Recent sediments (Obaje, 2009). The study area is one of the major production sites for tantalite, tin, niobium, columbite and gemstones, which are mined from mineralized pegmatite veins hosted within the quartz-mica schist and granite gneiss (Figure 3). These pegmatites occur as coarse-grained rocks and mainly trend in the NNW-SSE direction.

2.2. Sample collection and preparation

In this study, a total of 18 topsoil samples (0–20 cm) were randomly collected from the different parts of the study area. Composite samples were also collected in dug pits at a depth interval of 1 m, 2 m, and 3 m. These samples were carefully collected; freshly accumulated soils were avoided. A representative soil sample about 1 kg was subjected to pulverization, air-dried and later passed through a 200-mesh nylon sieve to remove unwanted materials. Proper delineation of the mining sites was done during the reconnaissance survey, which was followed by a systematic sampling pattern. The coordinates of the sample location were recorded with the aid of a GPS. All collected samples were stored in a clean polythene bags to avoid contamination before the laboratory analysis.

2.3. Sample analysis procedure and accuracy

The dried fine-grained soil particles were digested in aqua regia using a mixture of 25% HNO₃ and 75% HCl. After a near-total absorption by hydrofluoric-perchloric acid, the major and trace elements were analysed using the inductively coupled plasma-mass spectrophotometry (ICP-MS) in a certified commercial laboratory in Canada (Bureau Veritas Mineral Laboratories). The accuracy of the analytical methods was monitored by the repeated analysis of a standard reference material done together with a batch of soil samples according to the technique of Dulski (2001). These data give satisfactory results, with accuracy values within \pm 1–10% for various elements. The recovery rate for the analysis ranges from 95 to 110%. The minimum

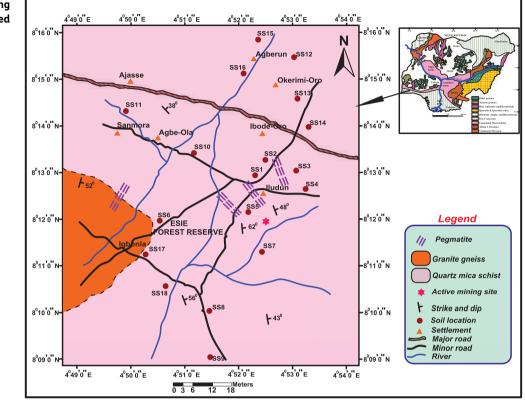


Figure 3. Location map showing sampling pattern superimposed on the geologic map.

detection limits for Fe, Pb, Zn, Cu, Cr, Ni, Ba, Sr, Th, Nb, Zr, Y, Sc and Mn were 0.04, 0.1, 1.0, 0.1, 0.1, 0.1, 0.1, 1.0, 0.5, 0.2, 0.1, 0.1, 0.1, 1.0 and 0.01, respectively.

2.4. Statistical analysis and spatial distribution maps

PCA was used to distinguish the different groups of heavy metals and classify heavy metals from different sources based on similarities in their chemistry. The location and spatial distribution maps of heavy metals were produced with the method of kriging interpolation technique using the software package SURFER version 15.0 (Golden Software Inc., CO, USA).

2.5. Methods of heavy metal pollution assessment

2.5.1. Contamination factor (CF)

CF is a quantitative evaluation of the level of contamination and sources of pollution. The CF makes use of conservative elements as the reference element (Cabrera, Clemente, Barrientos, López, & Murillo, 1999). CF is expressed as follows:

$$C_f^i = C_i / C_{ri} \tag{i}$$

where C_i is the content of metal *i* instead of mean content from at least five sample sites; C_{ri} is the reference value, baseline level of metal *i*. The degree of heavy metal pollution of CF is divided into four classes.

2.5.2. Geoaccumulation index (Igeo)

Igeo index as depicted by Eq. (ii) is used to classify identification of environmental influences by anthropogenic activities, also used to quantify the extent of heavy metal contamination associated with soils. Igeo is expressed as follows (Vural, 2015; Zhao, Feng, Yang, Niu, & Shen, 2012):

$$I_{geo} = \log_2 \frac{Cn}{1.5*Bn}$$
(ii)

where *Cn* is the measured concentration of metal *n*, *Bn* is the geochemical background value and constant 1.5 is used to represent the natural fluctuations and minor anthropogenic influences. The Igeo was distinguished into seven classes.

2.5.3. Contamination degree (CD)

This is expressed by the summation of all the CFs for each metal in the sequence. CD is expressed as follows:

$$CD = \sum_{(i=1)}^{n} Cf$$
 (iii)

where n is the number of metals, i is the baseline level of the metal and Cf is the CF. The CD is divided into four classes.

2.5.4. Pollution load index (PLI)

PLI is used to determine the magnitude of heavy metal contamination in soils. It is a potent tool in heavy metal pollution evaluation (Tomlinson, Wilson, Harris, & Jeffrey, 1980). The PLIs are calculated for all the metals. PLI is expressed as follows:

$$\mathsf{PLI} = (Cf_1 \times Cf_2 \times Cf_3 \dots \times Cf_n)^{1/n}$$
(iv)

where Cf_1 is the CF and *n* is the number of metals. The degree of heavy metal risk assessed by PLI is mainly divided into three classes (Chakravarty & Patgiri, 2009).

2.5.5. Ecological risk index (ERI)

The ERI is used to evaluate heavy metal pollution in soil to associate ecological and environmental effects with their toxicology and the toxic-response factor (T_{ri}) of Cu, Zn, Cd, Cr, Ni and Pb is 5, 1, 30, 2, 5 and 5 (µg/g), respectively (Hakanson, 1980). ERI is expressed as follows:

$$E_{r}^{i} = T_{r}^{i} \times Cf_{i} \tag{V}$$

where T_r^i is the toxic-response factor for a given substance and Cf_i is the CF. The degree of heavy metal risk assessed by ERI is mainly divided into five classes.

2.5.6. Potential ecological risk index (PERI)

PERI is employed to determine the semi-quantitative evaluation of regional pollution level. It can be expressed as follows (Wang et al., 2015):

$$PERI = \sum_{i=1}^{n} E_{r}^{i}$$
(vi)

where E_r^i is the ecological risk factor of element *i*. The degree of heavy metal assessment by PERI is categorized into four classes.

3. Results and discussion

3.1. Heavy metals concentration in soils

The statistics of heavy metals are presented in Table 1. Concentrations of Pb, Zn, Cu, Cr, Ni, Ba, Sr, Th, Zr, Y, Sc and Mn in the soils, as well as the background values of heavy and trace metals, are presented. The concentrations of Pb, Zn, Fe, Cu, Cr and Ni varied between 0–678, 7–222, 1.64–8.87, 8.8–145, 1.4–81 and 1.6–93.1 ppm, respectively, with average concentrations of 107.72, 48.17, 5.17, 45.28, 14.76 and 21.53 ppm, respectively. The order of abundance of these heavy metals analysed in the soils followed this pattern Pb > Zn> Cu> Ni> Cr> Fe. The values of Pb, Fe, Cu, Ba and Sr are higher than the background values (samples collected 3.5 km away from the mining site)

Table 1. Th	ne summary o	of chemical p	arameters of	f the soil sam	ples of the study	/ area
Elements	Minimum	Maximum	Mean	Standard deviation	Coefficients of variation (%)	Background value
Fe (%)	1.64	8.87	5.17	2.22	42.9	4.06
Pb (ppm)	0	678	107.72	174.12	161.64	19
Zn (ppm)	7	222	48.17	63.09	130.97	72
Cu (ppm)	8.8	145	45.28	39.01	86.15	29
Cr (ppm)	1.4	81	14.76	18.62	126.15	18
Ni (ppm)	1.6	93.1	21.53	27.31	126.85	24
Ba (ppm)	63	3022	626.19	916.57	146.37	573
Sr (ppm)	18.1	570.2	107.74	168.54	156.43	63
Th (ppm)	3.3	16.7	11.29	4.77	42.25	21
Nb (ppm)	23.76	972.05	119.33	215.41	180.52	125
Zr (ppm)	60.2	488.2	255.71	125.13	48.93	409
Y (ppm)	7.8	71.2	27.53	18.63	67.67	33
Sc (ppm)	6	36	15.33	9.09	59.30	13
Mn (ppm)	0.01	0.11	0.05	0.04	80	127.6

indicating that mining activities had influence in the study area. The concentrations of Pb, Cu, Ba, Sr, Nb and Zr varied greatly across the study area. Mn is relatively low and slightly homogenous in all the soil samples analysed. The coefficient of variation (CV) is a measure of relative variability and used to compare the degree of variation from one data series to another. Fe, Th and Zr had the lowest CV of 42.9, 42.25, and 48.93%, respectively, followed by Cu, Mn and Y having CVs of 86.15, 80 and 67.67%, respectively. Pb, Zn, Cr, Ni, Ba, Sr and Nb had their CV more than 90% (161.64, 130.97, 126.15, 146.37, 156.43 and 180.52%, respectively). This indicates that the (CV < 90%) values of heavy metals are dominated by anthropogenic sources are generally high compared to (CV > 90%) dominated by geogenic sources. Several other factors could be responsible for the large coefficient variation across the study area and these might include the following: sampling methods, sample preparation methods and analytical techniques. However, these variations could be further separated into direct and indirect contributions to the total variation with the use of PCA. Therefore, Pb, Zn, Cr, Ni, Ba, Sr and Nb concentrations in the soil samples tend to be high and introduced by anthropogenic activities whilst Fe. Th and Zr are associated with natural sources. The high concentration of Pb, Cu and Sr (Table 2) in the soisls posed a serious threat to vegetation and animal rearing and ultimately affecting the overall quality of human health through consumption. Moderately high concentrations of Fe, Zn, Cu and Cr indicate an evidence of leaching of mine waste under acidic conditions. In general, different sample points showed varied soil heavy metal concentrations, indicating influence of human activities from mineral exploitation and processing.

3.2. Spatial distribution of heavy metals

The spatial distribution of heavy metals concentrations is a veritable tool used to access the possible sources of enrichment and identify likely hotspot regions with high metal concentration (Li, Lu, Zhang, Liu, & Hou, 2008). In this study, the concentration variability maps of Zn, Pb, Ni, Mn, Fe, Cu, Cr and Sr were made with the method of kriging interpolation using the Surfer (version 15.0) software package. The spatial distribution maps of the study area are presented in Figure 4a-h. The distribution patterns of Zn, Ni, Mn, Fe, Cu, Cr and Sr were decreasing in concentrations away from the active mining site to the adjacent non-active mining region. They peaked close to the mining site.

The spatial distribution of Zn, Ni, Mn and Cu vary slightly from other metals. They peaked at two points located in the northwestern part of the study area, associated with increased agricultural

Table 2.	Geochemi	cal data o	Table 2. Geochemical data of selected elements in	elements		samples	the soil samples of the study area	dy area							
					Ele	ements (al	Elements (all the elements are in ppm except Fe in $\%$)	ients are ii	n ppm exc	ept Fe in 9	(%				
Sample ID	Ъþ	υZ	Cu	C	ïZ	Co	Ba	Sr	ТҺ	qN	Zr	7	Sc	Fe	ЧИ
SS1	24	24	32.2	12.3	12.3	0.1	232	71.5	16.7	92.39	488.2	28.1	19	6.01	0.09
SS2	34	14	20	8.4	8.3	BDL	166	42.4	16.1	138.42	278	13.3	11	6.43	0.02
SS3	26	129	98.5	21.5	55.7	0.1	656	30	10.1	127.86	195.5	794	26	7.13	0.08
SS4	71	∞	13.8	4.3	3.6	0.1	101	39.6	7.7	40.1	148.1	9.7	7	3.37	0.02
SS5	46	L	8.8	1.4	1.6	0.1	92	92.9	3.3	972.05	401	14.4	7	1.64	0.02
SS6	78	10	13.4	2.4	2.7	BDL	63	18.1	4.1	26.89	64.2	7.8	9	2.85	0.01
SS7	31	20	30.8	12.3	13.1	0.1	182	45.6	14.8	70.61	288.2	19.5	16	5.30	0.06
SS8	134	145	107.1	24.2	65.3	BDL	762	30.3	10.1	58.24	184.1	55.3	30	7.44	0.09
SS9	456	30	34.7	12	15.3	0.1	258	69	16	70.28	388.9	26.5	18	6.01	0.09
SS10	678	10	12.4	2.1	2.9	0.1	63	18.6	4.3	43.06	60.2	6'.	9	2.92	0.01
SS11	68	128	92.7	20.6	59.4	0.1	634	31.5	9.4	64.18	221.3	49	27	6.98	0.09
SS12	29	14	24	81	L.T	BDL	217	6''98	15	76.06	296.6	15.2	11	6.78	0.06
SS13	51	13	18	7.4	8.2	0.1	166	4.84	15.1	111.2	236.6	12.2	11	6.37	0.02
SS14	46	52	45.4	13.9	24.5	0.1	379	67.1	14	90.24	265.8	26.8	17	6.77	0.07
SS15	56	21	28.8	6.6	7.6	0.1	137	96.5	10.2	73.25	206.8	13.3	12	4.29	0.04
SS16	46	10	45.7	2.3	3.1	0.1	3022	570.2	15.3	34.32	397.6	39.7	8	1.88	0.02
SS17	12	10	43.8	2.4	3.1	0.1	3012.5	561.7	16.1	35.14	399.2	39.3	8	1.90	0.02
SS18	0	222	145	30.5	93.1	0.1	1129	21.1	5	23.76	82.4	71.2	36	8.87	0.11
*BDL = Belc	*BDL = Below Detection Limit	Limit													

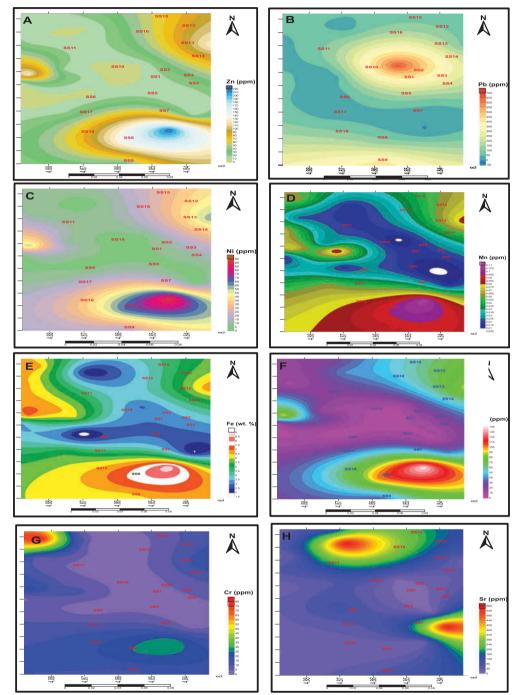


Figure 4. a-h: Spatial distribution maps of Zn, Pb, Ni, Mn, Fe, Cu, Cr and Sr in study area.

activities and high traffic density. The second peak is located in the southeastern part of the study area, related to the presence of active mining of tantalum-niobium concentrates. Pb distribution across the study area showed high spatial variability. Only one hotspot is visible, and it is located close to the central area of the study area around the mines. The spatial distribution of Fe is quite different from other heavy metals. A single peak around the mine with relatively low spatial variability is shown. The mean concentration is comparable to the background value, suggesting that Fe concentrations in the soils have not been altered by anthropogenic activities. The spatial distribution of Cr showed relatively low spatial variability across the study area. A single hotspot is

shown at the extreme northwestern region indicating the presence of intense agricultural activities. Cr may also be inputted into the soil by the parent material.

Also, the spatial distribution of Sr concentrations showed very high spatial variability with two distinct peaks around the mining area and the agricultural land space. The mean concentrations of Sr are higher compared with the background values, suggesting that Sr concentrations in the soils of the study area have been affected by anthropogenic activities or derived mainly due to the input of weathered parent material.

3.3. Indices of pollution assessment

3.3.1. Assessment using CF, CD, Igeo and PLI

Based on the monitoring data of soil quality in the study area, a quantitative analysis of heavy metal pollution assessment around the tantalum-niobium mining area was undertaken to determine the environmental effects. The evaluation was done using the CF and geoaccumulation index as presented in Tables 3 and 4.

The CF was calculated by comparing the observed heavy metal concentration with the background values for soil taken in an undisturbed location about 3.5 km away from the mining site. The CF values followed the descending sequence: Pb > Sr> Cu> Fe> Sc> Ba> Nb> Ni> Y> Cr> Zn> Zr> Th (Figure 5). Pb and Sr had the higher CF values of 35.68 ± 0.1 and 9.05 ± 0.05 , indicating that the soils were considerably contaminated to highly contaminated. The values of the CF of various trace and heavy metals in the soils are summarized in Table 3 is as follows; Zn (0.10–3.08), Cu (0.43–5.00), Cr (0.08–4.50), Ni (0.07–3.88), Ba (0.11–5.27), Th (0.16–0.80), Nb (0.19–1.11), Zr (0.15–1.19), Y (0.24–2.16), Sc (0.46–2.77) and Fe (0.40–2.18). The values derived from these metals indicate that their mode of occurrence was a natural process (geogenic). Zn, Cu, Cr, Ni and Ba showed low-to-considerable CF (Cf < 1: $3 \le Cf < 6$), whilst Th, Nb, Zr, Y, Sc and Fe showed low-to-moderate CF (Cf < 1: $1 \le Cf < 3$) as presented in Table 5.

The Igeo values for each heavy metals are Pb (-1.25 to 4.57), Zn (-3.32 to 1.62), Cu (-2.32 to 2.59), Cr (-4.32 to 1.58), Ni (-4.64 to 1,37), Ba (-3.84 to 1.82), Sr (-2.40 to 2.59), Th (-3.32 to -0.92), Nb (-2.94 to 2.37), Zr (-3.32 to -0.32), Y (-2.64 to 0.53), Sc (-1.69 to 0.89) and Fe (-1.69 to 0.55). The Igeo values followed the descending order: Pb > Cu> Fe> Sc> Sr> Y> Nb> Zr> Ba> Zn> Th> Cr> Ni (Figure 6). Most of the values indicate uncontaminated to moderately contaminated, except for Pb that indicates uncontaminated to heavily contaminated. Based on the average values, the soils of the study area are mostly slightly contaminated by almost all elements and more pronounced for Pb. Zn, Cu, Cr, Ni and Ba showed uncontaminated (Igeo < 0) to moderately contaminated (1 < Igeo < 2) of the class 1 order. Also, Th, Nb, Zr, Y, Sc and Fe showed uncontaminated (Igeo < 0) to moderately contaminated (1 < Igeo < 2) of the class 2 order (Table 5).

The degree of contamination, which is expressed by the summation of all the CFs for each metal, varies from moderate degree of contamination to very high degree of contamination; indicating an alarming anthropogenic influence. Zn, Cr, Ni, Ba, Nb, Y, Sc and Fe showed a considerable degree of contamination for all the soil samples, whilst Pb, Cu and Sr indicate a very high degree of contamination depicting alarming anthropogenic input due to mining activities and increased agricultural activities.

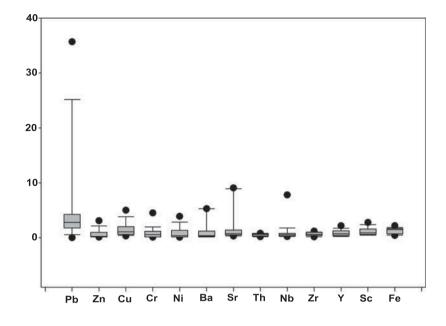
The values of the PLI in Table 5 were found to be generally (<1) in most of the sampled points except for three sites which is slightly higher than 1. The PLI provides a simple, comparative means to determine the magnitude of heavy metals in a site. A value of PLI = 0 indicates perfection, PLI = 1 indicates only baseline levels of pollutants, whilst PLI > 1 indicates progressive deterioration of site (Chakravarty & Patgiri, 2009).

	Fe	1.48	1.58	1.76	0.83	0.40	0.70	1.31	1.83	1.48	0.72	1.72	1.67	1.57	1.67	1.06	0.46	0.47	718
	_	1	1	1	0	0	0	1	1	1	0	1	1	1	1	1	0	0	(
	Sc	1.46	0.85	2.00	0.54	0.54	0.46	1.23	2.31	1.38	0.46	2.08	0.85	0.85	1.31	0.92	0.62	0.62	770
	۲	0.85	05.0	1.41	0.29	0.44	0.24	0.59	1.68	0.80	0.24	1.48	0.46	0.37	0.81	0,40	1.20	1.19	2 1 F
	Zr	1.19	0.68	0.48	0.36	0.98	0.16	0.70	0.45	0.95	0.15	0.54	0.73	0.58	0.65	0.51	0.97	0.98	0.00
	qN	0.74	1.11	1.02	0.32	7.78	0.22	0.56	0.47	0.56	0.34	0.51	0.61	0.89	0.72	0.59	0.27	0.28	010
	۲	0.80	0.77	0.48	0.37	0.16	0.20	0.70	0.48	0.76	0.20	0.45	0.71	0.72	0.67	0.49	0.73	0.77	7C U
	Sr	1.13	0.67	0.48	0.63	1.47	0.29	0.72	0.48	1.10	0.30	0.50	1.35	0.77	1.07	1.53	9.05	8.92	0 33
tudy area	Ba	0.40	0.29	1.14	0.18	0.16	0.11	0.32	1.33	0.45	0.11	1.11	0.38	0.29	0.66	0.24	5.27	5.26	1 97
les of the s	ï	0.51	0.35	2.32	0.15	0.07	0.11	0.55	2.72	0.64	0.12	2.48	0.32	0.34	1.02	0.32	0.13	0.13	3 88
e soil samp	ŗ	0.68	0.47	1.19	0.24	0.08	0.13	0.68	1.34	0.67	0.12	1.14	4.50	0.41	0.77	0.37	0.13	0.13	169
netals in th	Cu	1.11	0.69	3.40	0.48	0.30	0.46	1.06	3.69	1.20	0.43	3.20	0.83	0.62	1.57	0.99	1.58	1.51	5 00
factor for n	Zn	0.3	0.19	1.79	0.11	0.10	0.14	0.28	2.01	0.42	0.14	1.78	0.19	0.18	0.72	0.29	0.14	0.14	3.08
amination	Pb	1.26	1.79	2.95	3.74	1.79	4.11	1.63	7.05	24.00	35.68	4.68	3.53	2.68	1.79	2.95	1.79	0.63	0
Table 3. Contamination factor for metals in the soil samples of the study area	Sample ID	SS1	SS2	SS3	SS4	SS5	SS6	SS7	SS8	SS9	SS10	SS11	SS12	SS13	SS14	SS15	SS16	SS17	SC1.8

	Fe	-0.01	0.08	0.23	-0.86	-1.89	-1.09	-0.20	0.29	-0.01	-1.06	0.20	0.15	0.07	0.15	-0.49	-1.69	-1.69	0.55
	Sc	-0.04	-0.84	0.41	-1.47	-1.47	-1.69	-0.29	0.62	-0.12	-1.69	0.46	-0.84	-0.84	-0.20	-0.69	-1.29	-1.29	0.89
	۲	-0.81	-1.89	-0.09	-2.32	-1.79	-2.64	-1.36	0.16	-0.89	-2.64	-0.01	-1.69	-2.00	-0.89	-1.89	-0.32	-0.34	0.53
	Zr	-0.32	-1.15	-1.64	-2.06	-0.62	-3.32	-1.09	-1.74	-0.67	-3.32	-1.47	-1.06	-1.36	-1.22	-1.56	-0.62	-0.62	-2.94
	ЧN	-1.03	-0.43	-0.56	-2.25	2.37	-2.84	-1.36	-1.69	-1.43	-2.12	-1.56	-1.29	-0.76	-1.06	-1.36	-2.47	-2.40	-2.94
	Ę	-0.92	-0.97	-1.64	-2.06	-3.32	-2.94	-1.09	-1.64	-0.97	-2.84	-1.74	-1.06	-1.06	-1.18	-1.64	-1.03	-0.97	-2.64
8	Sr	-0.40	-1.15	-1.64	-1.25	-0.03	-2.40	-1.06	-1.64	-0.45	-2.32	-1.60	-0.15	-0.97	-0.49	0.03	2.59	2.57	-2.18
e study area	Ba	-1.89	-2.40	-0.40	-3.06	-3.18	-3.84	-2.25	-0.17	-1.74	-3.84	-0.43	-2.00	-2.40	-1.18	-2.64	1.82	1.81	0.39
soil samples of the study area	ï	-1.56	-2.12	0.63	-3.32	-4.64	-3.64	-1.47	0.86	-1.22	-3.64	0.72	-2.25	-2.12	-0.56	-2.25	-3.47	-3.47	1.37
	c	-1.12	-1.69	-0.32	-2.64	-4.32	-3.47	-1.12	-0.15	-1.18	-3.64	-0.40	1.58	-1.89	-0.97	-2.06	-3.47	-3.47	0.18
or metals in	Cr	-0.43	-1.12	1.18	-1.64	-2.32	-1.69	-0.49	1.30	-0.32	-1.79	1.09	-0.86	-1.29	0.06	-0.60	0.07	0.01	1.74
ion index fo	Zn	-1.60	-2.40	0.84	-3.18	-3.32	-2.84	-1.84	1.01	-1.25	-2.84	0.83	-2.40	-2.47	-0.47	-1.79	-2.84	-2.84	1.62
Table 4. Geoaccumulation index for metals in the	Pb	-0.25	0.25	0.97	1.32	0.25	1.45	0.12	2.23	4.00	4.57	1.64	1.23	0.84	0.25	0.97	0.25	-1.25	0.00
Table 4. Ge	Sample ID	SS1	SS2	SS3	SS4	SS5	SS6	SS7	SS8	SS9	SS10	SS11	SS12	SS13	SS14	SS15	SS16	SS17	SS18

le 4. 0	le 4. Geoaccumulation index for met	ation index	for metals i	als in the soil samples of the study area	imples of th	ne study are	Ş						
ple	q	Zn	C	۲	ï	Ba	Sr	Ŧ	ЧN	Zr	۲	Sc	
	-0.25	-1.60	-0.43	-1.12	-1.56	-1.89	-0.40	-0.92	-1.03	-0.32	-0.81	-0.04	
	0.25	-2.40	-1.12	-1.69	-2.12	-2.40	-1.15	-0.97	-0.43	-1.15	-1.89	-0.84	
	0.97	0.84	1.18	-0.32	0.63	-0.40	-1.64	-1.64	-0.56	-1.64	-0.09	0.41	
	1.32	-3.18	-1.64	-2.64	-3.32	-3.06	-1.25	-2.06	-2.25	-2.06	-2.32	-1.47	–
	0.25	-3.32	-2.32	-4.32	-4.64	-3.18	-0.03	-3.32	2.37	-0.62	-1.79	-1.47	ì
	1.45	-2.84	-1.69	-3.47	-3.64	-3.84	-2.40	-2.94	-2.84	-3.32	-2.64	-1.69	ì
	0.12	-1.84	-0.49	-1.12	-1.47	-2.25	-1.06	-1.09	-1.36	-1.09	-1.36	-0.29	-
	2.23	1.01	1.30	-0.15	0.86	-0.17	-1.64	-1.64	-1.69	-1.74	0.16	0.62	
	4.00	-1.25	-0.32	-1.18	-1.22	-1.74	-0.45	-0.97	-1.43	-0.67	-0.89	-0.12	
	4.57	-2.84	-1.79	-3.64	-3.64	-3.84	-2.32	-2.84	-2.12	-3.32	-2.64	-1.69	Ī
	1.64	0.83	1.09	-0.40	0.72	-0.43	-1.60	-1.74	-1.56	-1.47	-0.01	0.46	
	1.23	-2.40	-0.86	1.58	-2.25	-2.00	-0.15	-1.06	-1.29	-1.06	-1.69	-0.84	
_	0.84	-2.47	-1.29	-1.89	-2.12	-2.40	-0.97	-1.06	-0.76	-1.36	-2.00	-0.84	0
.+	0.25	-0.47	0.06	-0.97	-0.56	-1.18	-0.49	-1.18	-1.06	-1.22	-0.89	-0.20	
	0.97	-1.79	-0.60	-2.06	-2.25	-2.64	0.03	-1.64	-1.36	-1.56	-1.89	-0.69)
	0.25	-2.84	0.07	-3.47	-3.47	1.82	2.59	-1.03	-2.47	-0.62	-0.32	-1.29	ì
_	-1.25	-2.84	0.01	-3.47	-3.47	1.81	2.57	-0.97	-2.40	-0.62	-0.34	-1.29	ì
	0.00	1.62	1.74	0.18	1.37	0.39	-2.18	-2.64	-7,94	-7,94	0.53	0.89	

Figure 5. Indices of CF for metals in soil samples of the study area.



3.3.2. Assessment using potential ecological risk

The PERI was introduced to determine the semi-quantitative evaluation of regional pollution level (Wang et al., 2015), according to the toxicity of heavy metals and the response of the environment. Based on the standardized heavy metal toxic factor expressed by Hakanson (1980), the toxic response factor (T_{ri}) of Cu, Zn, Cd, Cr, Ni and Pb is 5, 1, 30, 2, 5 and 5 (µg/g), respectively. The ecological risk indices of Pb (0–178.4), Zn (0.1–3.08), Cu (1.5–18.45), Ni (0.6–13.6) and Cr (0.16–9.0) with an average of 28.35, 0.67, 7.81, 4.49 and 1.64, respectively, in all the sampling sites indicate low risk ($E_r^i < 30$). The highest ecological risk can be found in two sites with values (120, 178.4) indicating high risk. The main donation of the potential ecological risk comes from Pb, which indicates a significantly high risk with a value greater than 440.

3.4. Multivariate analysis

3.4.1. Pearson's correlation analysis

Pearson's correlation coefficient of heavy metals in the soil samples is presented in Table 6. There are varying levels of correlation between the metals in the soils. The inter-elemental relationship in the soils suggests different geochemical pathways are responsible for the metal concentration. A positive correlation between heavy metals suggests that these metals have common sources, mutual dependence and identical behaviour during transport. On the other hand, negative correlations suggest that they do not share the previously mentioned metal traits with each other. The results showed a strong correlation between Cu–Zn, Ni–Zn, Ni–Cu, Sr–Ba, Y–Cu, Sc–Zn, Sc–Cu, Sc–Ni, Sc–Y, Fe–Zn, Fe–Cu, Fe–Ni, Fe–Sc, Mn–Zn, Mn–Cu, Mn–Ni, Mn–Y, Mn–Sc and Mn–Fe. This may indicate that elements have similar sources of input or are precipitated in the same media. Pb, Th and Nb correlate negatively with other elements indicating that they originated from different or mixed sources. These diverse sources can be; gangue minerals, atmospheric deposition, community sewage system, use of pesticide and fertilizers amongst others.

3.4.2. Identification of source

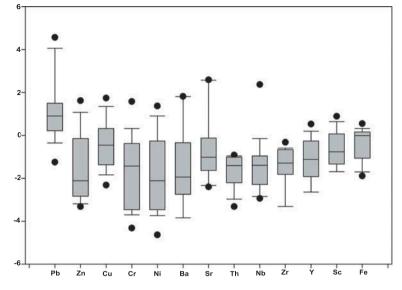
Using the multivariate statistical analysis, PCA can be used to identify the sources of pollution (Zheng-Qi, Shi-Jun, Xian-Guo, & Cheng-Jiang, 2008). The results of PCA for heavy metals concentration are presented in Table 7. Two rotated principal components (PCs) with eigenvalues of 6.401 and 2.859 accounted for 66.14% of the total variance (Figure 7). These results showed two component models for all the data variation. PC1 accounted for 45.72% of the

Table 5. Summary of	Table 5. Summary of the pollution indices calculated for the study area	ulated for the study area													
Indices	Equation	Classification	Ъb	Zn	Cu	ა	ïŻ	Ba	Sr	Ч	ЧN	Zr	۲	S	Fe
Contamination factor	$Cf = \frac{C_n}{B_n}$	Cf < 1 (low Cf)	2	14	∞	13	13	12	10	18	15	17	12	10	9
(Cf)		$1 \leq Cf < 3$ (moderate Cf)	6	m	9	4	4	4	9		2	1	9	∞	12
		3 ≤ Cf < 6 (considerable Cf)	4	1	4	1	1	2							
		Cf ≥ 6 (very high Cf)	m						2		1				
Geoaccumulation Index (Igeo)	$lgeo=log_2rac{Cn}{1.5*Bn}$	Igeo < 0 Class 0 (uncontaminated)	2	14	11	16	14	15	15	18	17	18	16	14	10
		0 < Igeo < 1 Class 1 (uncontaminated to moderately contaminated)	6	2	ſ		c	1	-				2	4	×
		1 < Igeo < 2 Class 2 (moderately contaminated)	4	2	4	-	1	2							
		2 < Igeo < 3 Class 3 (moderately to heavily contaminated)	1						2		1				
		3 < Igeo < 4 Class 4 (heavily contaminated)	7												
		4 < Igeo < 5 Class 5 (heavily to extremely contaminated)	1												
		Igeo ≥ 5 Class 6 (extremely contaminated)													
Contamination degree	$CD = \sum_{n}^{n} Cf$	CD < 6 (low CD)													
	5] []	$6 \leq CD < 12$ (moderate CD)	2												
		12 ≤ CD < 24 (considerable CD)	8												
		CD ≥ 24 (very high CD) indicating alarming anthropogenic contamination	m												
														(Cor	(Continued)

Indices	Equation	Classification	q	zn	C	Ⴆ	ī	Ba	JS	Th Nb	b Zr	~	ъ	Ъ.
Pollution Load Index	PLI =	PLI = 0 (perfection)	10			-	-	-	-	-	-	-	-	
(PLI)	$(Cf_1 \times Cf_2 \times Cf_3 \times Cf_n)^{1/n}$													
		PLI = 1 (baseline levels of pollutants present)												
		PLI > 1 (progressive deterioration of site)	m											
Ecological Risk (Er)	$E_r^i = T_r^i imes {\sf C} f_i$	E'_{r} < 30 (low risk)	15	18	18	18	18							
		E'_{r} : 30–60 (moderate risk)	-											
		E ¹ : 60–120 (considerable riśk)												
		E': 120–240 (high risk)	2											
		E' : > 240 (significantly high risk)												
Potential Ecological Risk (PERI)	$PERI = \sum_{i=1}^{n} E_{i}^{i}$	PERI < 110 (low risk)		7		-	7							
		PERI: 110–220 (moderate risk)			1									
		PERI: 220–440 (high risk)												
		PERI > 440 (significantly high risk)	I											

	Risk;	
	ological	ند
	Er = Ec	Elemen
	Index;	i of the
	on Load	entration
	= Polluti	d Conce
	e; PLI	groun
	Degree	= Back
	ation	n; Bn
	ontami	entratio
h	CD = C	nt Conc
_	ndex;	lemer
	tion I	rved E
	umula	= Obse
	eo-Acc	x; Cn
	eo = G	k Inde
_	or; Ig(cal Ris
	n Fact	cologi
	ninatic	ntial E
	ontam	: Potei

Figure 6. Indices of geoaccumulation for metals in soil samples of the study area.



total variance with strong loadings of Zn and Cu (0.781 and 0.873, respectively), supported by the strong correlations between Zn and Cu (Table 6) and their low Igeo and CF values. Hence, they may be contributed by geogenic sources. Coefficients of PC1 are supported by moderate loadings of Ni, Y, Sc, Fe and Mn (0.586, 0.635, 0.687, 0.629 and 0.643, respectively). These metals are siderophilic (having a strong affinity for metallic phase and usually depleted in the silicate portion of the earth) and were correlated well with one another.

The second principal component (PC2) accounted for 20.42% of the total variance with moderate loadings of Ba and Sr. This is further supported by weak correlation of Ba and Sr with other elements. Pb was equivocally different in the two-rotated component matrix; this is supported by the negative correlation of Pb with other elements. Pb average was 5.6-fold higher than the background value, indicating that it was inputted by anthropogenic activities. This was further supported by its high Igeo and CF values.

4. Conclusions

Different veritable tools, methods and indices were employed for the contamination assessment of soils heavy metals and trace elements in a typical active mining site of Iludun-Oro and its environs, southwestern Nigeria. This study was aimed at determining the spatial variances of heavy metals and identifying the pollution level, ecological risk level and potential sources of heavy metals in the soils. A total of 18 topsoil samples were randomly collected across different part of the study area and analysed using the ICP-MS. The concentrations of Pb, Zn, Fe, Cu, Cr and Ni varied between 0–678, 7–222, 1.64–8.87, 8.8–145, 1.4–81 and 1.6–93.1 ppm, respectively, with an average concentration of 107.72, 48.17, 5.17, 45.28. 14.76 and 21.53 ppm, respectively. The sequence of occurrence of these heavy metals analysed in the soils followed this pattern Pb > Zn> Cu> Ni> Cr> Fe. The values of Pb, Fe, Cu, Ba and Sr are higher than the background values indicating that mining activities influenced the study area. Pb, Zn, Cr, Ni, Ba, Sr and Nb had their CV more than 90% (161.64, 130.97, 126.15, 146.37, 156.43 and 180.52%, respectively). This indicates that the (CV < 90%) values of heavy metals are dominated by anthropogenic sources are high compared to (CV > 90%) dominated by geogenic sources.

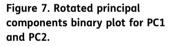
Results of the spatial distribution maps and multivariate statistical analysis of the heavy metals showed that mining activities, atmospheric deposition, poor community sewage system, and the use of pesticide and fertilizers represented the most important sources of heavy metal pollution. Accessing the soil quality in the study area, the results of the geoaccumulation, CF and PLIs

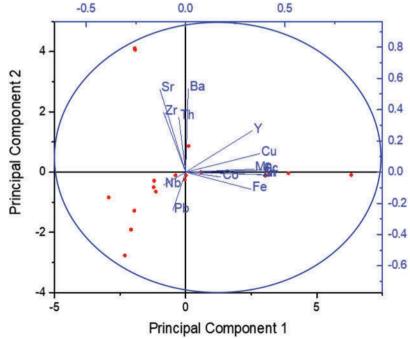
Table 6. R	esults of I	Pearson's d	Table 6. Results of Pearson's correlation matrix for soil samples	matrix for	soil sample	ŷ								
Element	Ъb	Zn	cu	ŗ	ï	Ba	Sr	μT	ЧN	Zr	٨	Sc	Fe	Mn
Pb	1	143	199	130	146	237	205	213	130	249	235	154	136	084
Zn			.970**	.308	**799.	960.	283	289	174	384	.872**	.938**	**969.	.753**
Cu			1	.293	.967**	.307	067	143	249	254	.954**	.922**	.631**	.746**
ت د				1	.325	126	215	.160	160	037	.218	.347	.567*	.470*
ïZ					1	.070	307	244	186	362	.867**	.956**	.733**	.780**
Ba						1	.920**	.280	205	.322	.523*	600.	303	077
Sr							1	.382	073	.507*	.181	333	554*	328
μ								1	343	.671**	004	039	.207	.142
qN									1	.326	204	212	323	205
Zr										1	012	168	193	.072
~											7	.845**	.489*	.708**
Sc												1	.816**	.910**
Fe													1	.795**
Mn														1
**. Correlation	n is significa	nt at the 0.0	**. Correlation is significant at the 0.01 level (2-tailed)	1).										

*. Correlation is significant at the 0.05 level (2-tailed).

Bold correlation is significant at the 0.05 level.

Table 7. The result of pri	ncipal component analy	ysis for soils	
	Comp	onent	
Component matrix	Coefficients of PC1	Coefficients of PC2	Communalities
Pb	-0.065	-0.248	0.066
Zn	0.781	-0.015	0.610
Cu	0.873	0.118	0.776
Cr	0.179	-0.032	0.046
Ni	0.586	-0.017	0.627
Ва	0.013	0.533	0.284
Sr	-0.129	0.528	0.295
Th	-0.034	0.352	0.125
Nb	-0.111	-0.084	0.019
Zr	-0.111	0.382	0.158
Υ	0.635	0.265	0.474
Sc	0.687	0.014	0.472
Fe	0.629	-0.110	0.506
Mn	0.643	0.019	0.432
Initial Eigenvalues	6.401	2.859	
Percentage of Variance (%)	45.72	20.42	
Cumulative percentage (%)	45.72	66.14	





indicated high concentrations of Pb in the soils of Iludun-Oro. Also, most of the values indicated uncontaminated to moderately contaminated for all elements except for Pb. This research suggested that the lessening of anthropogenic sources in the study area should be considered in further future control measures. Our study highpoints that multiple indices provide more reasonable assessments, which will be considered in the future studies on soils with heavy metals.

Conclusively, this paper has provided background information for a vulnerability assessment of the study area as well as for future impact assessment and proper planning of the mining site. Besides, this study would recommend a remediation plan for analysis of the quality of drinking water and soil quality for crop production in the study area. To sum up, there is the need for implementation and enforcement of environmental laws, which may be an aid to pollution control around the mining sites.

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Competing Interest

The authors declare no competing interests.

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Authors' contributions

All the authors contributed equally to the preparation of this manuscript. All authors read and approved the final manuscript.

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